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**Groundwater Monitoring Plan for
the Nonradioactive Dangerous
Waste Landfill**

J. W. Lindberg
M. J. Hartman

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August 1999



Prepared for the U.S. Department of Energy
under Contract DE-AC06-76RLO 1830

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Pacific Northwest National Laboratory
Richland, Washington 99352

Summary

The Nonradioactive Dangerous Waste Landfill (NRDWL), which received nonradioactive hazardous waste between 1975 and 1985 is regulated under the *Resource Conservation and Recovery Act of 1976* (RCRA) and monitored by Pacific Northwest National Laboratory. Monitoring is done under interim-status, indicator-evaluation requirements (WAC 173-303 and by reference, 40 CFR 265.92). The well network includes three upgradient and 6 downgradient wells. The wells are sampled semiannually for contaminant indicator parameters and site-specific parameters and annually for groundwater quality parameters.

Upgradient Wells	699-26-34A 699-26-35A (shared with Solid Waste Landfill) 699-26-35C
Downgradient Wells	699-25-33A 699-25-34A 699-25-34B 699-25-34D 699-26-33 699-26-34B
Groundwater Quality Parameters	Chloride Iron Manganese Phenols Sodium Sulfate
Contaminant Indicator Parameters	pH Specific Conductance Total Organic Carbon Total Organ Halogens
Site-Specific Parameters	Nitrate Volatile Chlorinated Hydrocarbons

The purpose of this plan is to describe an efficient groundwater monitoring program that is capable of determining whether waste disposal at NRDWL has impacted groundwater quality in the uppermost aquifer.

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1.0 Introduction

The Nonradioactive Dangerous Waste Landfill (NRDWL), which received nonradioactive hazardous waste between 1975 and 1985, is located in the central Hanford Site (Figure 1.1) in southeastern Washington State. The Solid Waste Landfill, which is regulated and monitored separately, is adjacent to the NRDWL. The NRDWL is regulated under the *Resource Conservation and Recovery Act of 1976* (RCRA) and monitored by Pacific Northwest National Laboratory.^(a) Monitoring is done under interim-status, indicator-evaluation requirements (WAC 173-303 and by reference, 40 CFR 265.92). The well network includes three upgradient wells (one shared with the Solid Waste Landfill) and six downgradient wells. The wells are sampled semiannually for contaminant indicator parameters and site-specific parameters and annually for groundwater quality parameters.

1.1 Purpose

The purpose of this plan is to describe a streamlined groundwater monitoring program that is capable of determining whether waste disposal at NRDWL has impacted groundwater quality in the uppermost aquifer. This document supersedes all previous monitoring plans (Weekes et al. 1987; Hodges 1993a; Hodges 1995). A revision to Hodges 1993a (Hodges 1995) incorporated the new wells suggested in Hodges monitoring program into the monitoring network and proposed an additional, deep well. Subsequent evaluation indicates that an additional deep well is not needed. The monitoring program proposed in this document is based on current conceptualization of the site and is consistent with data collected during 12 years of monitoring the site.

1.2 Regulatory Status and History

In November 1980, an initial RCRA part A permit application for NRDWL was submitted to the U.S. Environmental Protection Agency. The application was most recently revised in 1990, when a closure/postclosure plan also was submitted (DOE 1990). However, that plan was never approved or implemented.

In 1986, a groundwater monitoring program compliant with WAC 173-303-400(3) and 40 CFR 265, Subpart F (interim status) was required by a consent agreement and compliance order from the Washington State Department of Ecology. These requirements did not change under the *Hanford Federal Facility Agreement and Consent Order* (Tri-Party Agreement; Ecology et al. 1989). Nine wells were installed, seven of which comprised the initial monitoring network (Weekes et al. 1987). In 1987, quarterly sampling to establish background levels began. Sampling was reduced to a semiannual schedule in 1989 following four quarters of background data collection. Two new wells were installed in 1992. Interim-status indicator evaluation has provided no indication of significant groundwater contamination from NRDWL.

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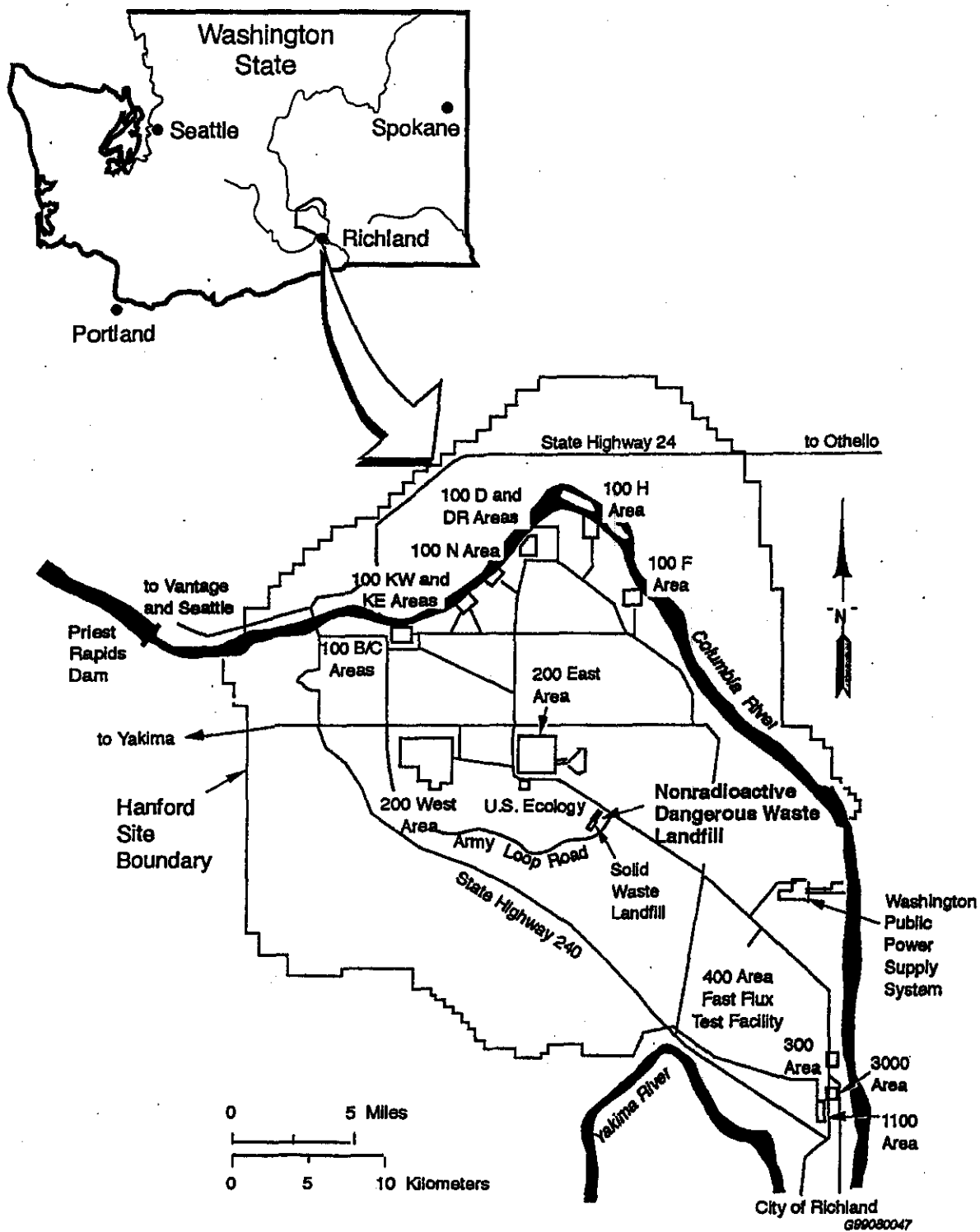


Figure 1.1. Location of the Nonradioactive Dangerous Waste Landfill

2.0 Description of the Nonradioactive Dangerous Waste Landfill

2.1 Physical Structure and Operational History

The NRDWL is located ~5.6-km southeast of the 200 East Area on the Hanford Site. The landfill has an area of 4.5 hectares and began operation in 1975. It consists of 19 parallel trenches, each ~122-m long, 4.9-m wide at the base, and 4.6-m deep. Beginning in 1975, chemical waste was disposed of in six trenches, asbestos in nine trenches, nonhazardous solid waste in one trench, and three were unused. The last receipt of dangerous waste was in May 1985, and the last receipt of asbestos occurred in May 1988. At the end of each operating day, the waste containers were covered with soil. This daily burial practice provided a temporary cover for the waste, but a permanent cover is planned for site closure.

The Solid Waste Landfill is adjacent to NRDWL on the south side. It is a larger facility (27 hectare) that received principally solid waste, including paper, construction debris, asbestos, and lunchroom waste. It also received sewage and bus garage wash water. Formerly both landfills were operated as a single landfill (Central Landfill).

2.2 Waste Types

The waste disposed of in NRDWL falls into the following categories (Hodges 1993b):

- Bulk organic waste: solvent waste, paints, paint thinners, and waste oils.
- Metal cleaner waste: primarily a mixture of sulfamic acid and sodium bisulfate.
- Small-quantity laboratory chemicals: used and unused reagents and various laboratory formulations, primarily metallic salts, acids, bases, oxidizers, and organic chemicals.
- Asbestos: primarily building demolition material, which accounts for more than 50% by volume of all waste disposed in the landfill.
- Nonhazardous solid waste: office and lunchroom waste, construction and demolition debris, and septic tank sludge.

Most of the chemical waste was placed in metal drums before disposal. Containers of small-quantity laboratory chemicals were placed in lab-packs and surrounded with sorbing material. Nonhazardous waste and asbestos were generally not placed in containers. In addition, some of the bulk organic wastes that were sorbed onto soil and other sorbents may not have been placed in containers.

3.0 Hydrogeology

The geology and hydrology of the NRDWL site are described in detail by Weekes et al. (1987) and Hodges (1993a). The following summary is taken largely from those documents unless indicated otherwise.

3.1 Physical Hydrogeology

The NRDWL is underlain by sands and gravels of the Hanford and Ringold formations (Figure 3.1). The vadose zone is ~40-m thick and consists of sand, silty sandy gravel, and gravel. The water table is near the top of a silty sand unit of the Hanford formation. Saturated sediments are composed of the following units:

- Saturated Hanford sediments: gravelly sand to sandy gravel, ~18-m thick; estimated horizontal hydraulic conductivity from field aquifer tests is approximately 1,000 m/d.
- Upper Ringold and Ringold Formation unit E, divided into three units based on lithology and hydraulic conductivity (40 to 45-m thick):
 - Slightly silty gravelly sand to sand, ~4-m thick; estimated horizontal hydraulic conductivity from field aquifer tests is 60 m/d.
 - Hard, clayey silt (low permeability) 1 to 4-m thick; estimated vertical hydraulic conductivity from field aquifer tests ranges from 0.006 to 3 m/d.
 - Silty sand to sandy gravel, unknown thickness; estimated horizontal hydraulic conductivity from field aquifer tests ranges from 0.3 to 15 m/d. This unit is probably unit E, but there are no wells in the vicinity that fully penetrate this unit. Approximately 2-km east at well 699-25-26, the Ringold Formation unit E is 40 m thick (Lindsey 1991).
- Ringold Formation unit C, unit B, the lower mud unit, and unit A are described as follows:
 - Unit C, 10 m, gravel and sandy gravel
 - Overbank deposit, 20 m, sandy silt and silty sand
 - Unit B, 10 m, gravel and sandy gravel
 - Lower mud unit, 17 m, silt and sandy silt
 - Unit A, 28 m, gravel and sandy gravel
- Top of basalt at ~185-m depth

The uppermost aquifer is unconfined and is part of the saturated Hanford sediments and probably the upper portion of the Upper Ringold unit. A low-permeability unit perhaps in the lower portion of the Upper Ringold unit or in unit E is believed to form the base of the uppermost aquifer because the hydraulic conductivity of the aquifer base is orders of magnitude lower than the overlying sediments.

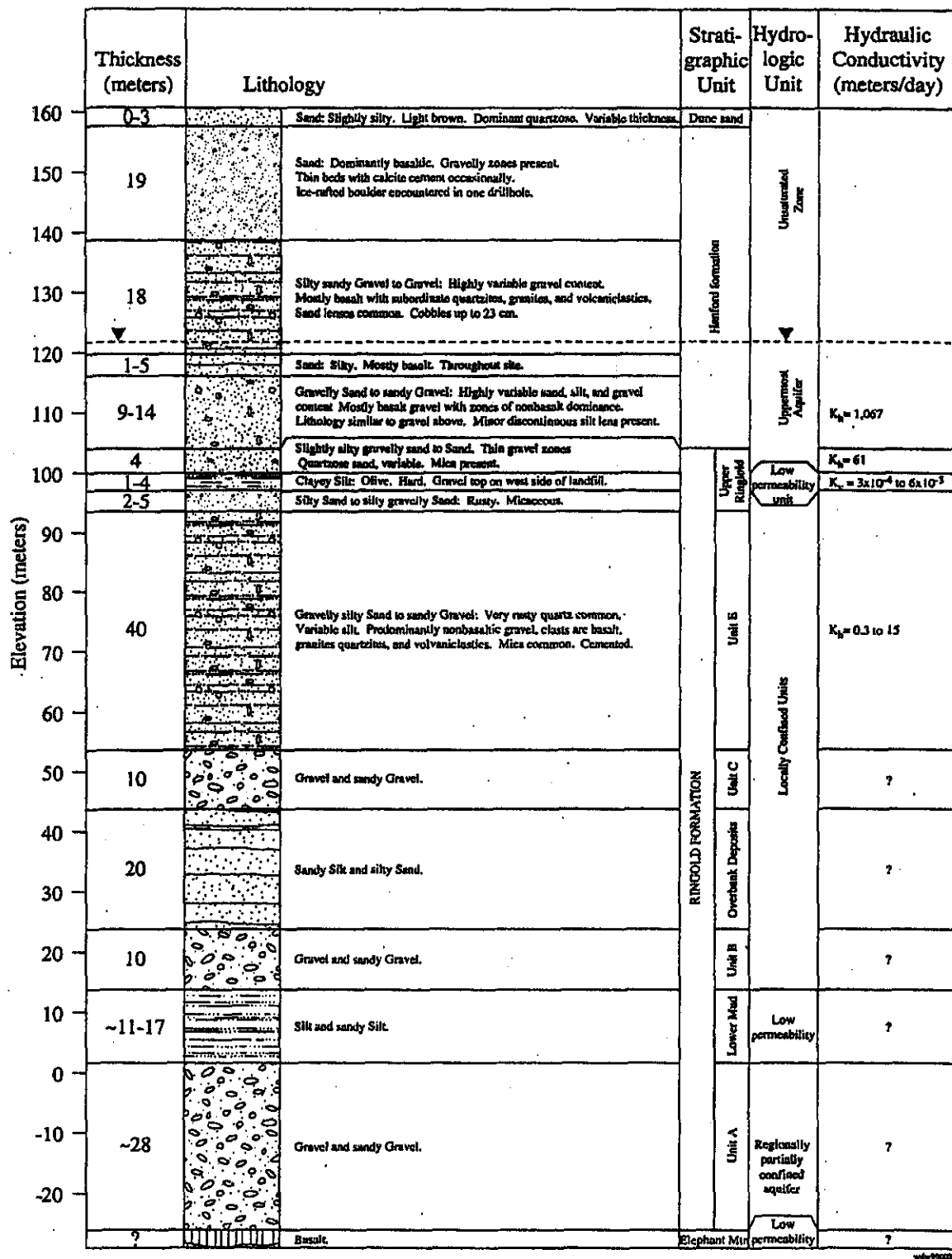


Figure 3.1. Stratigraphic Column at the NRDWL and Associated Hydrogeology

The hydraulic gradient beneath NRDWL is very low (0.00005; Weekes et al. 1987) because the aquifer is very transmissive. Previous estimates based on data from the nearby wells indicate that flow is generally west to east (Weekes et al. 1987, p. 43). A water-table map of the region around the landfill is shown in Figure 3.2. Assuming groundwater flows perpendicular to the equipotential lines, flow converges from the north-northeast and the southwest, and moves toward the southeast. Contaminant plumes originating in the 200 East Area move through the area from northwest to southeast, corroborate this interpretation.

Two wells at NRDWL sample the bottom of the uppermost aquifer, i.e., just above the low-permeability unit. Heads in these wells are virtually the same as in adjacent wells completed at the top of the aquifer, indicating no significant vertical gradient (Lindberg 1999, Section 3.7).

Water-levels beneath NRDWL declined nearly 2 m since 1988 because lower volumes of liquid waste are being discharged to the ground in the 200 Areas (Figure 3.3) (Lindberg 1999, Section 3.7). Based on a comparison of present levels of the water table with a hindcast water-table map (estimating water-table elevations in 1944, ERDA 1975) the water table could decline as much as 4.6 m before returning to pre-Hanford Site levels.

3.2 Summary of Groundwater Monitoring Results

Monitoring began at NRDWL and adjacent Solid Waste Landfill in 1987. Wells were sampled quarterly in 1987 through 1989, and semiannually thereafter. Some of the wells are co-sampled with Hanford environmental surveillance monitoring.

Concentrations of RCRA indicator parameters (specific conductance, pH, total organic carbon, and total organic halogens) have not significantly increased (or pH decreased) over background (upgradient) concentrations. Some chlorinated hydrocarbons were detected in the NRDWL monitoring wells in concentrations below their maximum contaminant levels (Table 3.1). One potential source of these low levels of chlorinated hydrocarbons is vadose-zone transport from the adjacent Solid Waste Landfill. Chlorinated hydrocarbons were disposed at the Solid Waste Landfill and are detected in groundwater downgradient of the Solid Waste Landfill. Soil gas surveys at NRDWL have detected several volatile organic compounds including chlorinated hydrocarbons. However, the shallow nature of soil gas surveys to date makes it inappropriate to link chlorinated hydrocarbons disposed at the Solid Waste Landfill with the low levels of chlorinated hydrocarbons detected in the groundwater (Jacques and Kerkow 1993).

Vadose zone gases were sampled during installation of wells 699-26-33 and 699-25-34A in 1992 (Hodges 1993a). A chlorinated hydrocarbon, probably carbon tetrachloride, was detected as deep as 37 m, near the water table (a malfunctioning gas chromatograph prevented unique identification of the compound). A shallow vadose zone soil gas survey was conducted in 1993 (Hodges 1994). The survey found widespread acetone and several chlorinated hydrocarbons, most notably tetrachloroethylene and trichloroethylene. The highest concentrations were detected over the older chemical trenches near the east end of NRDWL. Carbon tetrachloride, 1,1,1-trichloroethane, and chloroform were also detected locally. The presence of chlorinated hydrocarbons in the vadose zone suggests the possibility of their migration from the NRDWL to groundwater, and in fact most of them have been detected in groundwater.

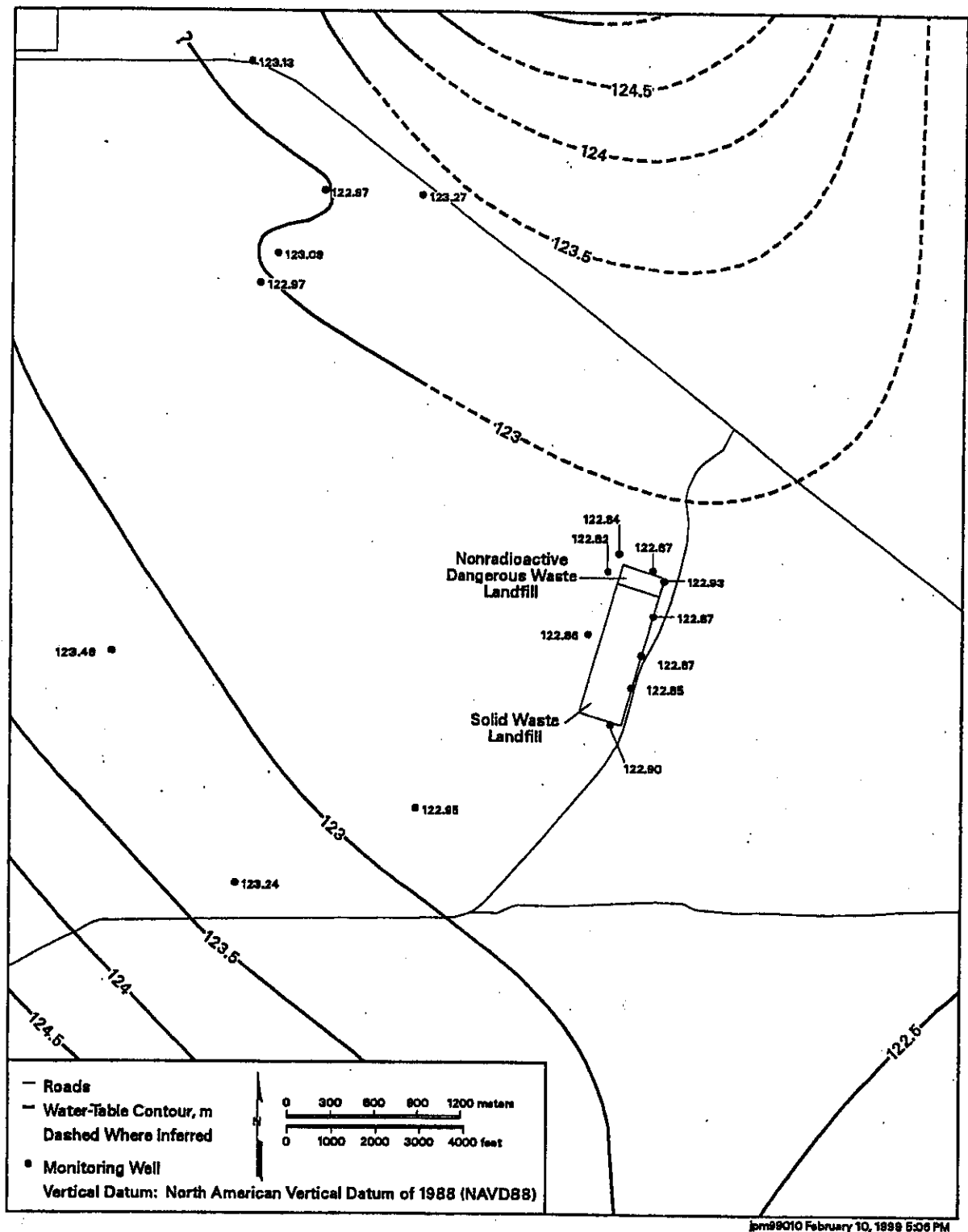


Figure 3.2. Water-Table Map for Nonradioactive Dangerous Waste Landfill and Solid Waste Landfill, June 1998

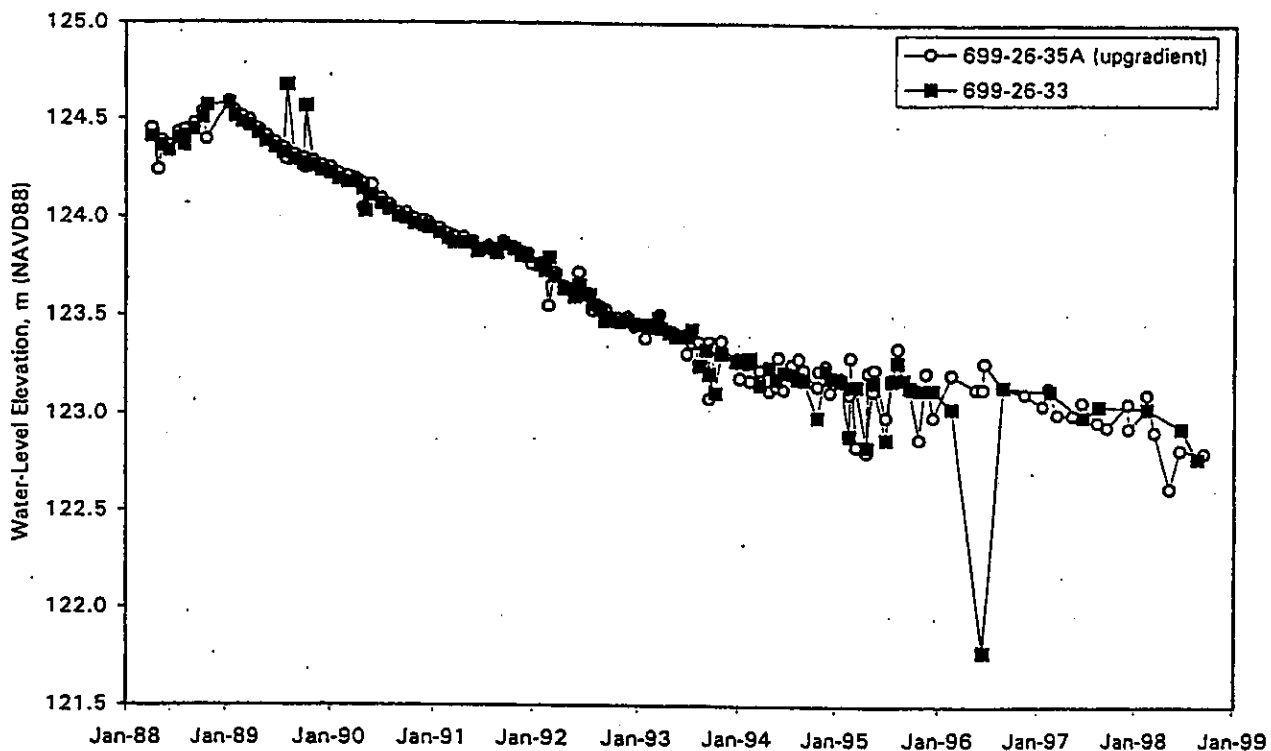


Figure 3.3. Water-Table Elevations in Wells Monitoring Nonradioactive Dangerous Waste Landfill

Tetrachloroethylene concentrations are higher in downgradient wells 699-25-34A and 699-25-34B than in upgradient wells (Figure 3.4). Downgradient concentrations are fairly steady at 1 to 2 $\mu\text{g/L}$. Trichloroethylene concentrations are also slightly higher in downgradient wells (Figure 3.5), but concentrations are less than or equal to 1 $\mu\text{g/L}$.

Carbon tetrachloride and 1,1,1-trichloroethane were detected in upgradient and downgradient NRDWL wells. All of the concentrations were less than or equal to 2 $\mu\text{g/L}$, and most are less than 1 $\mu\text{g/L}$ (Figures 3.6 and 3.7). Concentrations in the early 1990s were greater in downgradient wells 699-25-34A and 699-25-34B than in upgradient wells. Since then, concentrations have decreased and are approximately the same as in upgradient wells.

Chloroform was detected in downgradient wells. Recently concentrations have increased from below detection limits to 1 $\mu\text{g/L}$ in upgradient wells, which is greater than in downgradient wells (Figure 3.8). The cause of this change in upgradient concentrations is not known.

As mentioned previously in this section, acetone was also detected in shallow vadose zone gases. One set of samples from the NRDWL wells was analyzed for acetone in 1990. All results were below detection limits.

Groundwater beneath the NRDWL is contaminated with tritium, iodine-129, and nitrate from the 200 Areas. The plume boundaries bisect the Central Landfill, with low concentrations to the southwest and high concentrations to the northeast. Concentrations of these constituents in groundwater are decreasing gradually with time in all of the shallow NRDWL wells.

Table 3.1. Range and Average Concentration of Detected Chlorinated Hydrocarbons in NRDWL Wells, 1987-1998^(a)

Well	Constituent	Minimum µg/L	Maximum µg/L	Average ^(b) µg/L	Number of Sample Dates
699-26-34A upgradient	1,1,1-Trichloroethane	<DL	2.10	0.85	20
	1,1-Dichloroethane	<DL	0.14	0.01	20
	1,4-Dichlorobenzene	<DL	0.09	0.01	20
	Carbon tetrachloride	<DL	0.91	0.12	20
	Chloroform	<DL	1.0	0.07	20
	Tetrachloroethylene	<DL	0.70	0.26	20
	Trichloroethylene	<DL	0.46	0.11	19
	TOC	<DL	635.85	169.11	14
	TOX	<DL	3.75	0.83	22
699-26-35A upgradient	1,1,1-Trichloroethane	<DL	4.0	1.25	36
	1,1-Dichloroethane	<DL	0.30	0.02	36
	1,4-Dichlorobenzene	<DL	0.05	<DL	32
	Carbon tetrachloride	<DL	2.0	0.19	37
	Chloroform	<DL	1.0	0.06	37
	Tetrachloroethylene	<DL	3.0	0.46	36
	Trichloroethylene	<DL	1.0	0.16	35
	TOC	<DL	933	95	43
	TOX	<DL	9.8	2.1	32
699-25-34A downgradient	1,1,1-Trichloroethane	<DL	6.1	2.04	21
	1,1-Dichloroethane	<DL	0.3	0.05	21
	1,4-Dichlorobenzene	<DL	0.06	<DL	17
	Carbon tetrachloride	<DL	1.6	0.25	22
	Chloroform	<DL	0.5	0.15	22
	Tetrachloroethylene	<DL	1.5	0.77	21
	Trichloroethylene	<DL	0.7	0.28	20
	TOC	<DL	770	90	28
	TOX	<DL	10.7	2.4	23
699-25-34B downgradient	1,1,1-Trichloroethane	<DL	7.0	1.72	21
	1,1-Dichloroethane	<DL	0.3	0.06	21
	1,4-Dichlorobenzene	<DL	0.28	0.03	17
	Carbon tetrachloride	<DL	0.89	0.14	22
	Chloroform	<DL	1.1	0.16	22
	Tetrachloroethylene	<DL	1.75	0.69	21
	Trichloroethylene	<DL	1.1	0.3	20
	TOC	<DL	800	89	28
	TOX	<DL	16.2	3.6	23
699-25-34D downgradient	1,1,1-Trichloroethane	0.7	6.0	2.1	13
	1,1-Dichloroethane	<DL	2.0	0.22	13
	1,4-Dichlorobenzene	<DL	0.4	0.03	13
	Carbon tetrachloride	<DL	0.89	0.26	14

Table 3.1. (contd)

Well	Constituent	Minimum µg/L	Maximum µg/L	Average ^(b) µg/L	Number of Sample Dates
	Chloroform	<DL	0.33	0.13	14
	Tetrachloroethylene	<DL	2.0	0.87	13
	Trichloroethylene	<DL	0.90	0.48	12
	TOC	<DL	575	131	14
	TOX	<DL	11.6	4.8	9
699-26-33 downgradient	1,1,1-Trichloroethane	<DL	2.6	1.05	21
	1,1-Dichloroethane	<DL	0.13	0.01	21
	1,4-Dichlorobenzene	<DL	<DL	<DL	17
	Carbon tetrachloride	<DL	1.9	0.22	21
	Chloroform	<DL	0.30	0.06	21
	Tetrachloroethylene	<DL	0.92	0.40	21
	Trichloroethylene	<DL	0.52	0.14	20
	TOC	<DL	1470	160	28
	TOX	<DL	13	2.2	23
699-26-34B downgradient	1,1,1-Trichloroethane	0.24	1.4	0.99	13
	1,1-Dichloroethane	<DL	0.07	0.01	13
	1,4-Dichlorobenzene	<DL	<DL	<DL	13
	Carbon tetrachloride	<DL	0.60	0.18	13
	Chloroform	<DL	0.20	0.05	13
	Tetrachloroethylene	<DL	0.82	0.35	13
	Trichloroethylene	<DL	0.34	0.13	12
	TOC	<DL	636	169	14
	TOX	<DL	8.0	3.3	9
699-26-35C upgradient, deep	1,1,1-Trichloroethane	<DL	<DL	<DL	21
	1,1-Dichloroethane	<DL	<DL	<DL	21
	1,4-Dichlorobenzene	<DL	<DL	<DL	17
	Carbon tetrachloride	<DL	<DL	<DL	21
	Chloroform	<DL	0.20	0.02	21
	Tetrachloroethylene	<DL	<DL	<DL	21
	Trichloroethylene	<DL	0.07	<DL	20
	TOC	<DL	203	32	27
	TOX	<DL	9.8	1.2	21
699-25-33A downgradient, deep	1,1,1-Trichloroethane	<DL	0.25	0.02	21
	1,1-Dichloroethane	<DL	<DL	<DL	21
	1,4-Dichlorobenzene	<DL	<DL	<DL	17
	Carbon tetrachloride	<DL	<DL	<DL	22
	Chloroform	<DL	<DL	<DL	22
	Tetrachloroethylene	<DL	0.06	<DL	21
	Trichloroethylene	<DL	<DL	<DL	20
	TOC	<DL	1320	126	27
	TOX	<DL	6.4	0.87	22
(a) Excluded data flagged as suspect or rejected; averaged replicates by date.					
(b) Changed less-than detection values to zero to calculate average.					
DL = detection level					
TOC = total organic carbon.					
TOX = total organic halogen.					

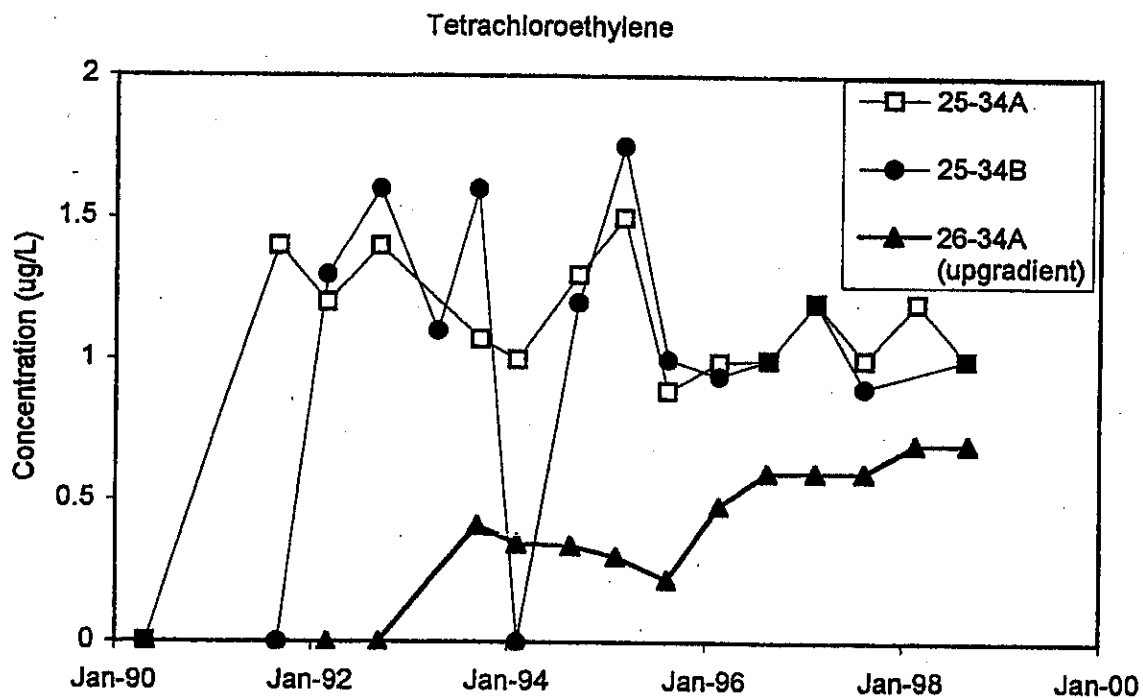


Figure 3.4. Tetrachloroethene Concentrations in NRDWL Wells

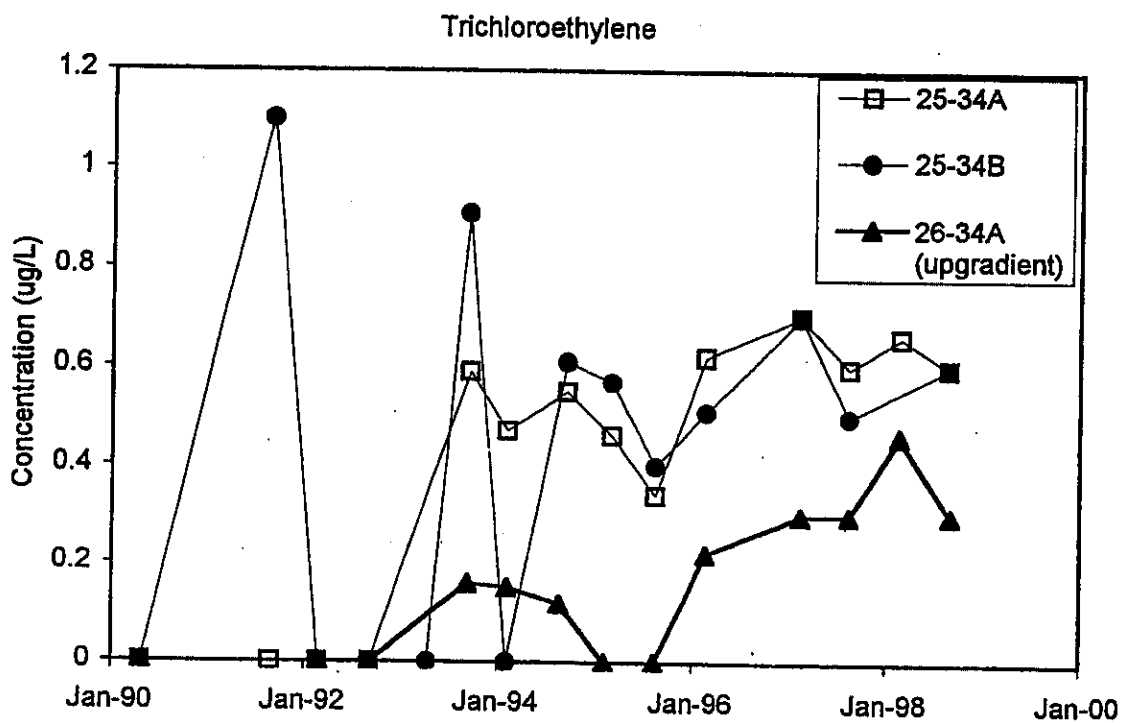


Figure 3.5. Trichloroethene Concentrations in NRDWL Wells

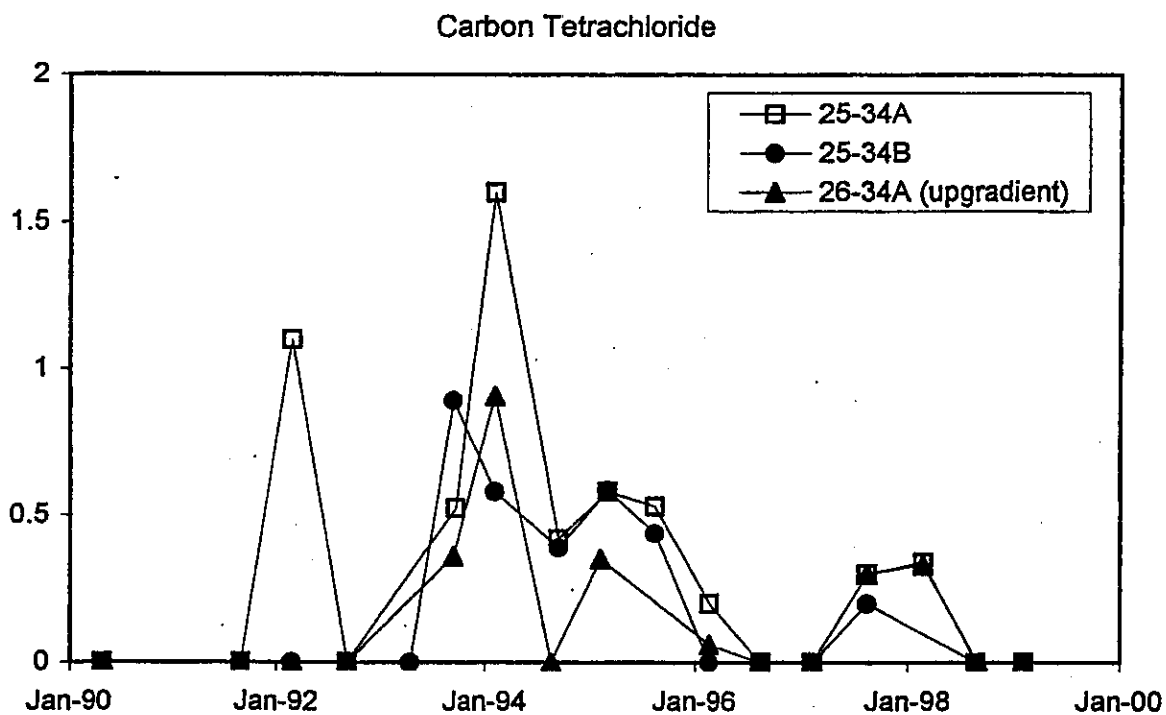


Figure 3.6. Carbon Tetrachloride Concentrations at NRDWL Wells

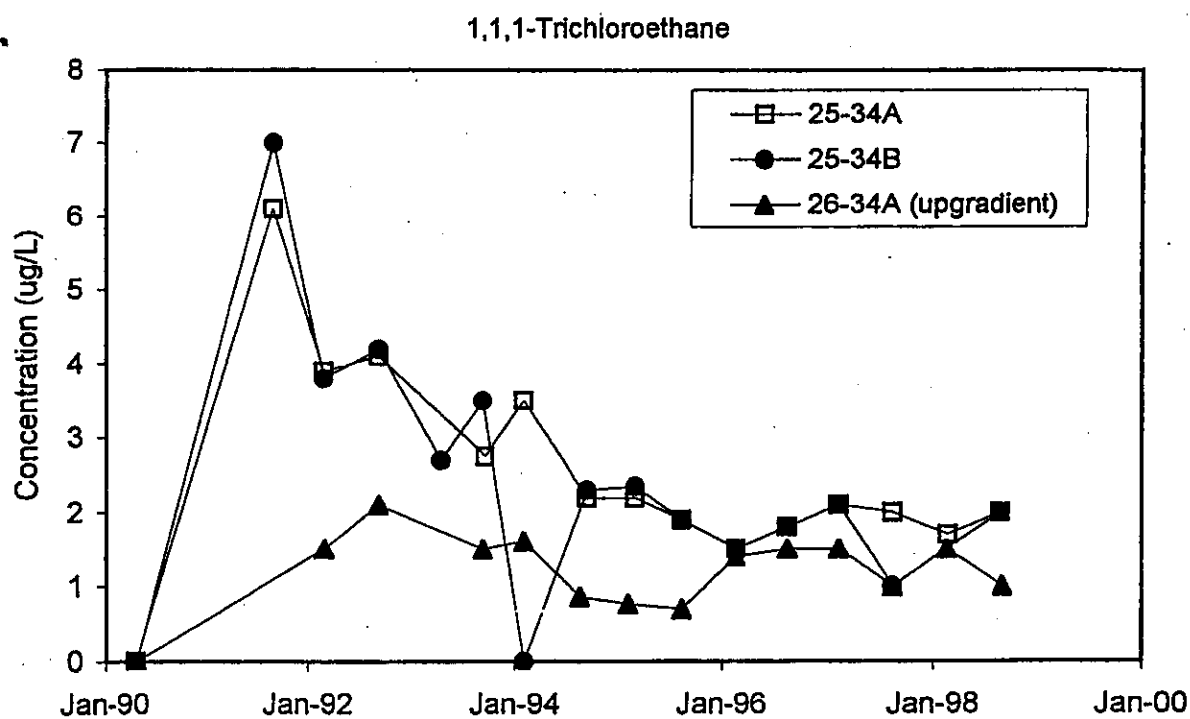


Figure 3.7. 1,1,1-Trichloroethane Concentrations at NRDWL Wells

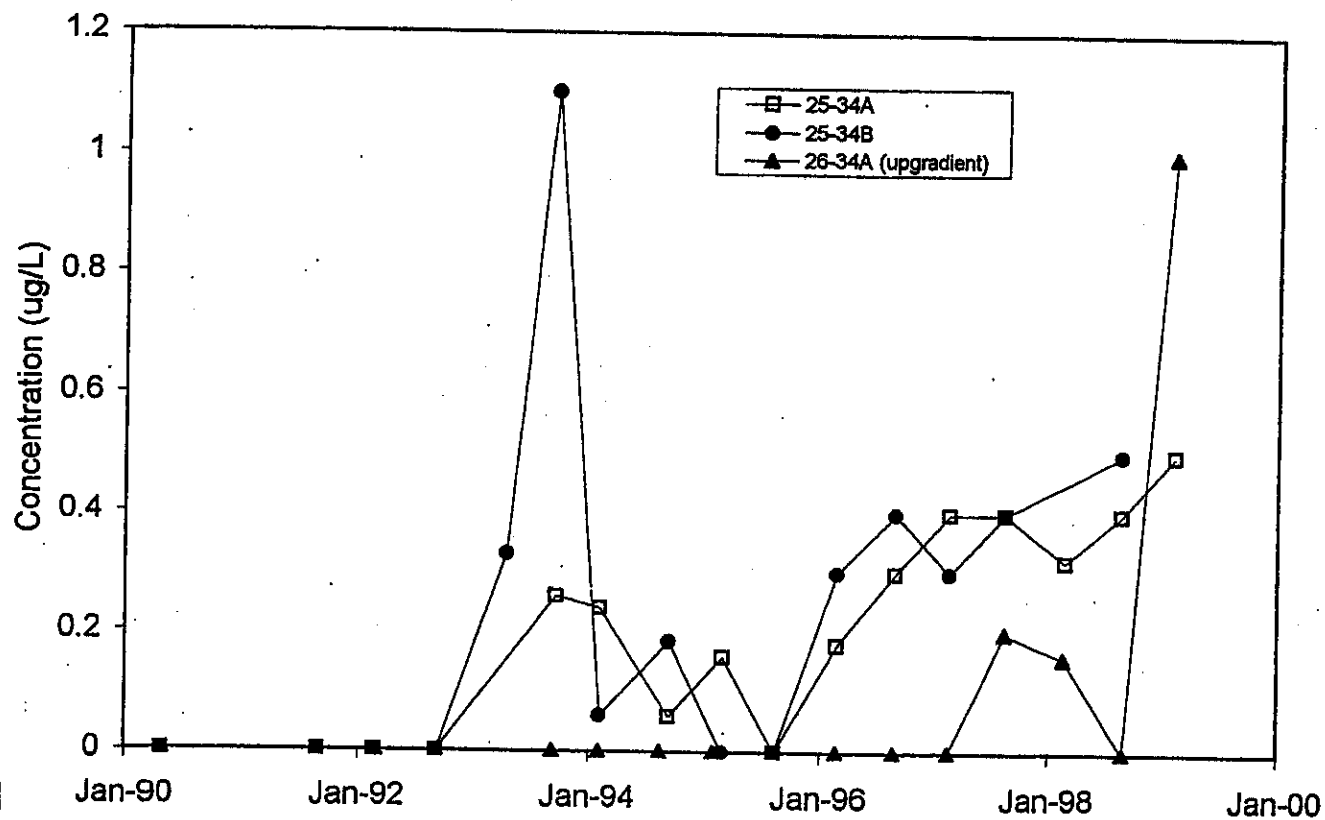


Figure 3.8. Chloroform Concentrations in NRDWL Wells

4.0 Conceptual Model

The following characteristics constitute key portions of the NRDWL conceptual model:

- Relatively small quantities of dangerous waste liquids were disposed of in NRDWL; most were placed in sorbing material so not much free liquid remained. It is unlikely that dense, nonaqueous phase liquids are present.
- Chlorinated hydrocarbons are present in vadose vapors and are the primary contaminant of concern for groundwater.
- Natural precipitation may carry some contaminants through the vadose zone to groundwater.
- Dangerous waste materials remain in the landfill; therefore, contamination from the landfill may still impact groundwater.
- Contaminants may move laterally within the vadose zone via vapor transport, perhaps from the adjacent solid waste landfill. However, soil gas studies to date have failed to prove this.
- Contaminants remain in the uppermost aquifer above the low-permeability unit. To date, indicator parameters measured in downgradient wells at the water table are lower in concentration than critical means or are within critical ranges, and contaminant concentrations are below respective MCLs. Similarly, the downgradient well sampling groundwater at the top of the low-permeability unit (well 699-25-33A) has very low concentrations of constituents monitored.
- The zone below the low-permeability unit has not been impacted by NRDWL because the saturated zone above the unit has not been impacted adversely by the NRDWL. That is, in order for the lower zone to be affected, the upper one must be affected first.
- Regionally, groundwater flows toward the southeast; however, flow directly beneath to the landfill may be toward the east or even northeast. The hydraulic gradient in the immediate vicinity of NRDWL is extremely low.

5.0 Groundwater Monitoring Program

5.1 Objectives of RCRA Monitoring

The overall objectives of the Groundwater Monitoring Program at the Hanford Site are to (a) protect human health and the environment; (b) comply with governmental regulations; and (c) contribute to groundwater investigation or remediation. Specifically, the objective of the groundwater monitoring at NRDWL is to detect adverse impact from the facility on the groundwater quality.

5.2 Special Conditions

Two hydrogeological conditions at NRDWL are of special concern to the development of this groundwater monitoring plan. The first is the low-permeability unit within Upper Ringold unit or Ringold Formation unit E. This low-permeability layer limits the thickness of the uppermost aquifer locally to about 22 m. It also limits the depth of contaminant sinkers (e.g., dense, nonaqueous phase liquids). A groundwater monitoring plan must account for this low-permeability zone and provide assurance that groundwater contamination from NRDWL has not reached the top of the low-permeability unit, as well as more shallow depths of the uppermost aquifer. (See also Section 4.0, "Conceptual Model.")

The second special condition is the extremely low hydraulic gradient and the difficulty in determining an accurate direction of groundwater flow in the uppermost aquifer. Water-table maps (like in Weekes et al. 1987, p. 43) indicate the flow should be generally from west to east in the immediate vicinity of the NRDWL. However, contaminate plumes like tritium from the 200 East Area are moving from the northwest to the southeast. (See also Section 4.0, "Conceptual Model.")

5.3 Sampling and Analysis Plan

5.3.1 Monitoring Well Network

The six downgradient wells and three upgradient wells (Table 5.1) of the monitoring well network are designed to

- detect groundwater contamination (from NRDWL) before it moves downgradient of the network wells
- compare upgradient and downgradient concentrations of indicator parameters
- determine if groundwater contamination has migrated vertically and impacted groundwater at the base of the uppermost aquifer (immediately above the low-permeability unit).

Table 5.1. Monitoring Well Network

Well	Year Installed	Hydrogeologic Unit Monitored	Well Construction Standard	Upgradient/Downgradient
699-25-33A	1987	Top of LPU ^(a)	WAC 173-160	Downgradient
699-25-34A	1986	Top of Unconfined Aquifer	WAC 173-160	Downgradient
699-25-34B	1986	Top of Unconfined Aquifer	WAC 173-160	Downgradient
699-25-34D	1992	Top of Unconfined Aquifer	WAC 173-160	Downgradient
699-26-33	1986	Top of Unconfined Aquifer	WAC 173-160	Downgradient
699-26-34A	1986	Top of Unconfined Aquifer	WAC 173-160	Upgradient
699-26-34B	1992	Top of Unconfined Aquifer	WAC 173-160	Downgradient
699-26-35A ^(b)	1986	Top of Unconfined Aquifer	WAC 173-160	Upgradient
699-26-35C	1987	Top of LPU ^(a)	WAC 173-160	Upgradient
(a) Low-permeability unit in Upper Ringold Formation.				
(b) Well shared with Solid Waste Landfill Network.				

The six downgradient wells are located (Figure 5.1) around the northern, eastern, and southern boundaries of NRDWL to detect potentially contaminated groundwater in response to groundwater flowing either eastward (interpreted from water table contours) or southeastward (interpreted from plume maps). Monitoring Efficiency Model (MEMO) results demonstrate that the downgradient wells have a monitoring efficiency of more than 90% for flow directions from 80 degrees to 140 degrees clockwise from north (DOE 1990). Results for a flow direction of 125 degrees clockwise from due north (southeast—the most likely flow direction) indicate a model efficiency of 99.8%. One downgradient well (699-25-33A) was installed at the top of the low-permeability unit to detect potentially contaminated groundwater at the base of the uppermost aquifer.

The three upgradient wells (Figure 5.1) are located northwestward to determine background water quality. Wells 699-26-34A and 699-26-35A are screened near the water table and are compared with downgradient water quality (from five downgradient wells) to determine if NRDWL has adversely affected groundwater quality. (See Sections 6.2 and 6.3 for an explanation of the methods used to compare background and downgradient water quality. Well 699-26-35A is shared with the Solid Waste Landfill.) Well 699-26-35C is screened immediately above the low-permeability zone, and results are used for information purposes only. Results from this well cannot be used for background statistics because the well monitors a different portion of the aquifer.

To determine whether groundwater contamination can be detected lower in the aquifer (lower than near the water table), two deeper wells sample groundwater at the top of the low-permeability zone. One of the deeper wells is located upgradient of NRDWL and the other is downgradient to examine upgradient and downgradient concentrations of chlorinated hydrocarbons over time (Figure 5.1). Some chlorinated hydrocarbons have a dense nonaqueous phase that could migrate downward to the top of the low-permeability zone.

Appendix A contains construction details for each well in the groundwater monitoring network at NRDWL.

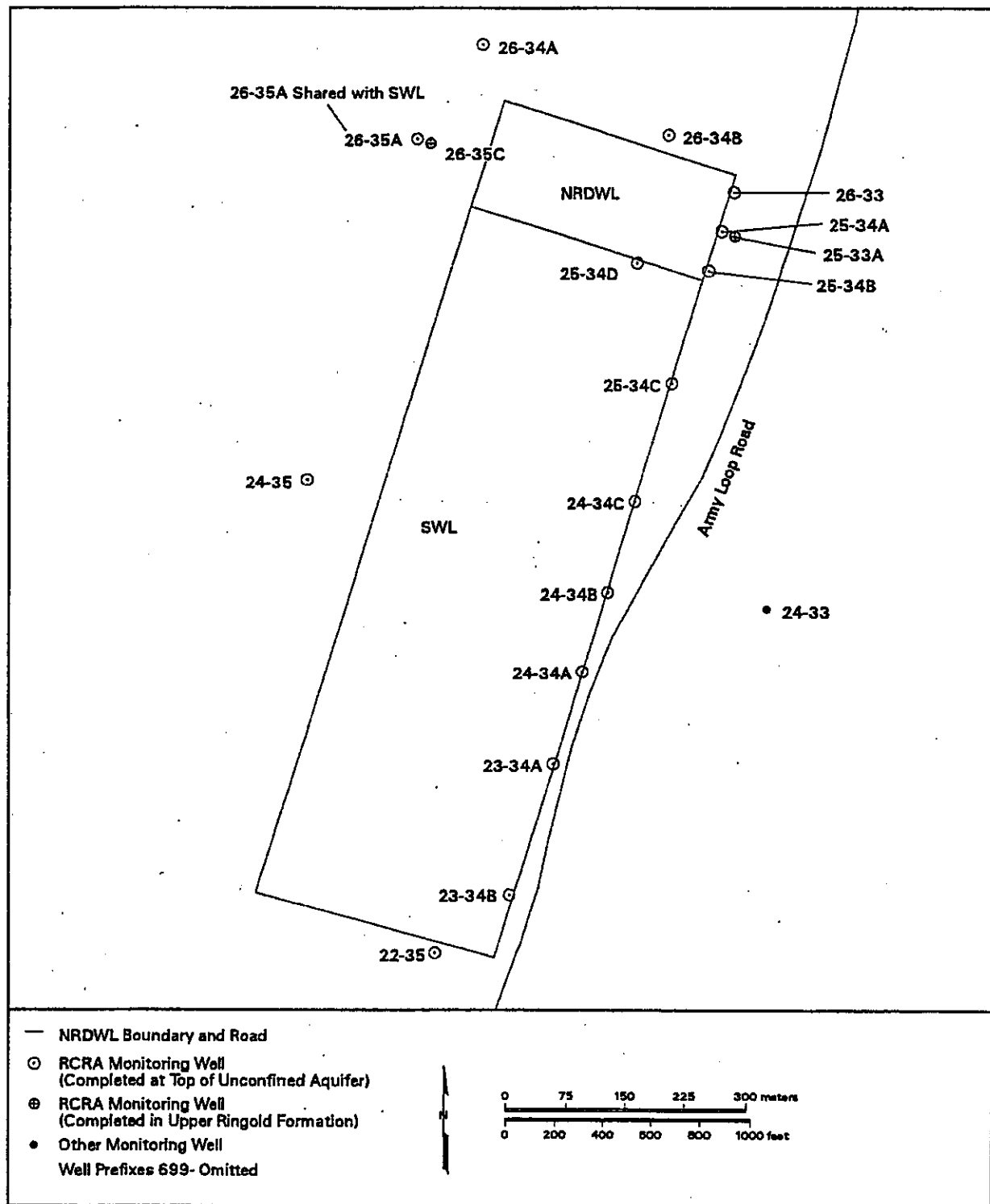


Figure 5.1. Monitoring Well Locations for Solid Waste Landfill and Nonradioactive Dangerous Waste Landfill

5.3.2 Constituent List and Sampling Frequency

After the first year, groundwater beneath RCRA treatment, storage, and disposal units in an interim status/indicator-evaluation program (WAC 173-303 and by reference, 40 CFR 265) must be monitored for groundwater quality parameters and contaminant indicator parameters. If appropriate, site-specific parameters may be added (Table 5.2 lists the monitored constituents and frequencies appropriate for NRDWL). The groundwater quality parameters are to be monitored annually, the contaminant indicator parameters semiannually, and the site-specific parameters semiannually. Groundwater quality parameters include chloride, iron, manganese, phenols, sodium, and sulfate. Contaminant indicator parameters include specific conductance, pH, total organic carbon, and total organic halogens. Appropriate site-specific parameters are nitrate and volatile chlorinated hydrocarbons. Nitrate is monitored semiannually because it is a significant upgradient groundwater contaminant that may affect an indicator parameter (specific conductance). Volatile chlorinated hydrocarbons are monitored semiannually because they have been detected in the 12 years of groundwater monitoring at NRDWL.

5.3.3 Determination of Groundwater Flow

Depth-to-water measurements will continue to be collected from each monitoring well when each is sampled. Therefore, depth-to-water will be measured at least semiannually at all network wells. These depth-to-water measurements will be used to construct water-table maps which, in turn, will be reevaluated annually to determine the direction of groundwater flow beneath NRDWL. In addition, depth-to-water will be measured at approximately six of the NRDWL monitoring wells during the month of March to support the efforts to make a water-table map of the entire Hanford Site for *Hanford Site Groundwater Monitoring* (Hartman 1999) each year. (Initially, the six wells measured will include 699-22-35, -23-34A, -25-34C, -26-34A, -26-34B, and -26-35A, but these may change as the needs of the Hanford Site map change.)

Table 5.2. Constituent List and Sampling Frequency

Contamination Indicator Parameters (Quadruplicate Samples) (Sampled Semiannually)
pH
Specific Conductance
Total Organic Carbon
Total Organic Halides
Groundwater Quality Parameters (Sampled Annually)
Chloride
Iron
Manganese
Phenols
Sodium
Sulfate
Site-Specific Parameters (Sampled Semiannually)
Volatile Chlorinated Hydrocarbons
Nitrate

The water table at NRDWL has a very low gradient. Water-table maps constructed previously from the depth-to-water measurements (and surveyed elevation of the ground surface and casing top at each well) show that the gradient across the site is approximately 0.00005 (Weekes et al. 1987). Determining groundwater flow direction in an area the size of the NRDWL when the gradient is 0.00005 is very difficult. Very small errors in depth-to-water measurements or in surveyed casing elevations are significantly large compared to the low gradient. Therefore, maps showing the major plumes of tritium, nitrate, and iodine-129 will be used to corroborate flow directions based on water-table maps.

Using the Darcy equation,

$$v = \frac{K(i)}{n_e} \quad (1)$$

average groundwater flow rate (v) will be estimated from known estimates of hydraulic conductivity (K), the water-table gradient (i), and effective porosity (n_e).

5.3.4 Sampling and Analysis Protocol

Monitoring at NRDWL is part of the Hanford Groundwater Monitoring Project. Procedures for groundwater sampling, documentation, sample preservation, shipment, and chain-of-custody requirements are described in Pacific Northwest National Laboratory or subcontractor manuals (currently, Waste Management Northwest procedures manual ES-SSPM-001, Waste Management Federal Services, 1998) and in the groundwater monitoring quality assurance (QA) plan (PNNL 1998). Samples generally are collected after three casing volumes of water have been purged from the well or after field parameters (pH, temperature, specific conductance, and turbidity) have stabilized. For routine groundwater samples, preservatives are added to the collection bottles before their use in the field. Samples to be analyzed for metals are usually filtered in the field so that results represent dissolved metals.

Procedures for field measurements are specified in the subcontractor's or manufacturer's manuals. Analytical methods are specified in contracts with laboratories, and most are standard methods from *Test Methods for Evaluating Solid Wastes, Physical/Chemical Methods* (EPA 1986b). Analytical methods are described in Gillespie (1999). Alternative procedures meet the guidelines of SW-846, Chapter 10 (EPA 1979).

5.3.5 Quality Assurance and Quality Control

The QA and quality control (QC) program for groundwater monitoring is designed to assess and enhance the reliability and validity of groundwater data. The primary quantitative measures or parameters used to assess data quality are accuracy, precision, completeness, and the method detection limit. Qualitative measures include representativeness and comparability. Goals for data representativeness for groundwater monitoring projects are addressed qualitatively by the specification of well locations, well construction, sampling intervals, and sampling and analysis techniques in the groundwater monitoring plan for each RCRA facility. Comparability is the confidence with which one data set can be compared

to another. The QC parameters are evaluated through laboratory checks (e.g., matrix spikes, laboratory blanks), replicate sampling and analysis, analysis of blind standards and blanks, and interlaboratory comparisons. Acceptance criteria have been established for each of these parameters (PNNL 1998), based on guidance from the U.S. Environmental Protection Agency (EPA 1986b). When a parameter is outside the criteria, corrective actions are taken to prevent a future occurrence and affected data are flagged in the database.

6.0 Data Management, Evaluation, and Reporting

This section describes how groundwater data are stored, retrieved, evaluated, and interpreted. Statistical evaluation methods and reporting requirements are also described.

6.1 Data Management

The contract laboratories report analytical results electronically. The results are loaded into the Hanford Environmental Information System (HEIS) database. Field-measured parameters are entered manually or through electronic transfer. Data from HEIS may be downloaded to smaller databases, such as the Geosciences Data Analysis Toolkit (GeoDAT) for data validation, reduction, and trend analysis. Paper data reports and field records are considered to be the record copies and are stored at Pacific Northwest National Laboratory.

The data undergo a validation and verification process according to a documented procedure, as described in the project QA plan. QC data are evaluated against the criteria listed in the project QA plan and data flags are assigned when appropriate. In addition, data are screened by scientists familiar with the hydrogeology of the unit, compared to historical trends or spatial patterns, and flagged if they are not representative. Other checks on data may include comparison of general parameters to their specific counterparts (e.g., conductivity to ions; gross alpha to uranium), calculation of charge balances, and comparison of calculated versus measured conductivity. If necessary, the laboratory may be asked to check calculations or reanalyze the sample, or the well may be resampled.

6.2 Interpretation

After data are validated and verified, the acceptable data are used to interpret groundwater conditions at the site. Interpretive techniques include the following:

- Hydrographs: graph water levels versus time to determine decreases, increases, seasonal, or man-made fluctuations in groundwater levels.
- Water-table maps: use water-table elevations from multiple wells to construct contour maps to estimate flow directions. Groundwater flow is assumed to be perpendicular to lines of equal potential.
- Trend plots: graph concentrations of chemical or radiological constituents versus time to determine fluctuations and trends. These plots may be used in tandem with hydrographs and/or water-table maps to determine if concentrations relate to changes in water-level or in groundwater flow directions.

- Plume maps: map distributions of chemical or radiological constituents areally in the aquifer to determine extent of contamination. Changes in plume distribution over time aid in determining movement of plumes and direction of flow.

6.3 Statistical Evaluation

The goal of RCRA detection monitoring is to determine if NRDWL has affected groundwater quality. This is determined based on the results of a statistical test. According to 40 CFR 265.92 [and by reference of WAC 173-303-400(3)], the owner/operator of an interim-status hazardous waste facility must establish initial background concentrations for the contamination indicator parameters: specific conductance, pH, total organic carbon, and total organic halogen. Four replicate measurements for each parameter from each well were collected at NRDWL quarterly for 1 year. Data from the upgradient wells were used to determine the initial background arithmetic mean and variance.

Monitoring data collected after the first year are compared with the initial background data to determine if there is an indication that contamination may have occurred. A t-test is required to make this determination [40 CFR 265.93(b)]. A recommended method is the averaged replicate t-test method described in Appendix B of the *RCRA Groundwater Monitoring Technical Enforcement Guidance Document* (EPA 1986a). The averaged replicate t-test method for each contamination indicator parameter is calculated as:

$$t = (\bar{x}_i - \bar{x}_b) / S_b * \sqrt{1 + 1/n_b}$$

where t = test statistic
 \bar{x}_i = average of replicates from the i^{th} monitoring well
 \bar{x}_b = background average
 S_b = background standard deviation
 n_b = number of background replicate averages.

A test statistic larger than the Bonferroni critical value, t_c , (i.e., $t > t_c$) indicates a statistically significant probability of contamination. These Bonferroni critical values depend on the overall false-positive rate required for each sampling period (i.e., 1% for interim status), the total number of wells in the monitoring network, and the number of degrees of freedom ($n_b - 1$) associated with the background standard deviation. Because of the nature of the test statistic in the above equation, results to be compared to background do not contribute to the estimate of the variance. The test can be reformulated, without prior knowledge of the results of the sample to be compared to background (i.e., \bar{x}_i), in such a way that a critical mean, CM, can be obtained:

$$CM = \bar{x}_b + t_c * S_b * \sqrt{1 + 1/n_b} \quad (\text{one tailed})$$

$$CM = \bar{x}_b \pm t_c * S_b * \sqrt{1 + 1/n_b} \quad (\text{two tailed})$$

If downgradient data exceed the CM, the data are determined to be statistically different from background. For pH, a two-tailed CM (or critical range) is calculated and downgradient data beyond the range are considered to be statistically different from background. If a statistical exceedance is detected, the well will be resampled to determine if the originally detected increase (or pH decrease) was a result of laboratory or measurement error (verification sampling). If verification sampling confirms the exceedance, the owner/operator must notify the Washington State Department of Ecology within 7 days and submit a groundwater quality assessment plan within 15 days following the notification [40 CFR 265.93(d)]. The assessment monitoring program determines if dangerous waste or dangerous waste constituents from the facility have entered the groundwater, their concentration, and their rate and extent of migration [40 CFR 265.93(d)]. Critical mean values for the NRDWL are presented in Table 6.1.

6.4 Reporting

Chemistry and water-level data are reviewed at least quarterly and are available in HEIS. Interpretive reports are issued annually in March (e.g., Hartman 1999). Reporting requirements are listed in Table 6.2.

Table 6.1. Critical Means for 28 Comparisons—Background Contamination Indicator Parameter Data for Nonradioactive Dangerous Waste Landfill^(a)

Constituent	n	df	t _c	Average Background	Standard Deviation	Critical Mean	Upgradient/Downgradient Comparison Value
Specific conductance, $\mu\text{S}/\text{cm}$	10	9	5.0255	461.65	4.845	487.2	487.2
Field pH	11	10	5.2814	7.45	0.164	[6.55, 8.35]	[6.55, 8.35]
Total organic carbon, ^(b) $\mu\text{g}/\text{L}$	11	10	4.8092	380.68	242.28	1,597.7	1,597.7
Total organic halides, ^(b,c) $\mu\text{g}/\text{L}$	11	10	4.8092	4.28	2.05	14.6	22.4
<p>(a) Data collected from August 1997 to February 1999 for upgradient wells 699-26-34A and 699-26-35A. (b) Critical means calculated from values reported below the contractually required detection limit. (c) The upgradient/downgradient comparison value is the limit of quantitation. df = Degrees of freedom (n-1). n = Number of background replicate averages. t_c = Bonferroni critical t-value for appropriate df and 28 comparisons.</p>							

Table 6.2. Reports Required for Compliance with 40 CFR 265, Subpart F, for Groundwater Monitoring

Submittal	Submittal Period	Reporting Vehicle	Regulatory Requirement ^(a)
First year of sampling: concentrations of interim primary drinking water constituents, identifying those that exceed limits.	Quarterly	Complete ^(b)	40 CFR 265.94(a)(2)(i)
Concentration and statistical analyses of groundwater contamination indicator parameters, noting significant differences in upgradient wells.	Annually, by March 1 of following year.	Hanford Groundwater Monitoring Report (e.g., Hartman 1999 ^(c))	40 CFR 265.94(a)(2)(ii)
Results of groundwater surface elevation evaluation and description of response if appropriate.	Annually, by March 1 of following year.	Hanford Groundwater Monitoring Report	40 CFR 265.94(a)(2)(iii)
Outline for groundwater quality assessment program	Within 1 year after effective date of regulations	Chapter 7 of this document	40 CFR 265.93(a)
Notification of statistical exceedance ^(d)	Within 7 days of verification	Letter to Ecology	40 CFR 265.93(c)
Assessment Plan ^(d)	Within 15 days of notification	PNNL document or letter	40 CFR 265.93(d)
Determinations under assessment program ^(d)	As soon as technically feasible; annually thereafter	PNNL document, letter, or Hanford Groundwater Monitoring Report	40 CFR 265.93(d)(5) and 265.94(b)
<p>(a) 40 CFR 265, U.S. Environmental Protection Agency. "Interim status standards for owners and operators of hazardous waste treatment, storage, and disposal facilities."</p> <p>(b) Requirement was fulfilled during first year of sampling via published reports. Quarterly submittal of data continues via HEIS.</p> <p>(c) Hartman MJ, ed. 1999. <i>Hanford Site groundwater monitoring for fiscal year 1998</i>. PNNL-12086, Pacific Northwest National Laboratory, Richland, Washington.</p> <p>(d) Required if exceedance occurs and is verified.</p>			

7.0 Assessment Monitoring

This section outlines the assessment monitoring plan for NRDWL, as required by 40 CFR 265.93(a). The assessment program must be capable of determining whether dangerous waste or dangerous waste constituents have entered the groundwater, their concentration, and the rate and extent of migration.

If an indicator parameter at a downgradient well significantly exceeds the background value, an assessment plan will be prepared and submitted to the Washington State Department of Ecology (see Section 6.3). The plan will include the following:

- the approach to determine if dangerous waste or dangerous waste constituents from the facility have entered the groundwater or if the exceedance was caused by other sources (false-positive rationale)
- the investigative approach to fully characterize the rate and extent of contaminant migration
- number, locations, and depths of wells in the monitoring network
- sampling and analytical methods used
- data evaluation procedures
- an implementation schedule.

An outline for the assessment plan is presented in Table 7.1.

As the assessment determinations are made, a report of the findings will be sent to the Washington State Department of Ecology. The determinations will then be updated annually as required by 40 CFR 265.94(b).

Table 7.1. Outline for Groundwater Quality Assessment Monitoring Plan

Introduction
Existing Data and Evaluation
Groundwater Quality Assessment Program
Approach
Assessment Monitoring Network
Constituents
Sampling and Analysis
Data Evaluation
Schedule
References

8.0 References

40 CFR 265, U.S. Environmental Protection Agency. "Interim status standards for owners and operators of hazardous waste treatment, storage, and disposal facilities."

DOE - U.S. Department of Energy. 1990. *Nonradioactive dangerous waste landfill closure/postclosure plan*. DOE/RL-90-17, U.S. Department of Energy, Richland, Washington.

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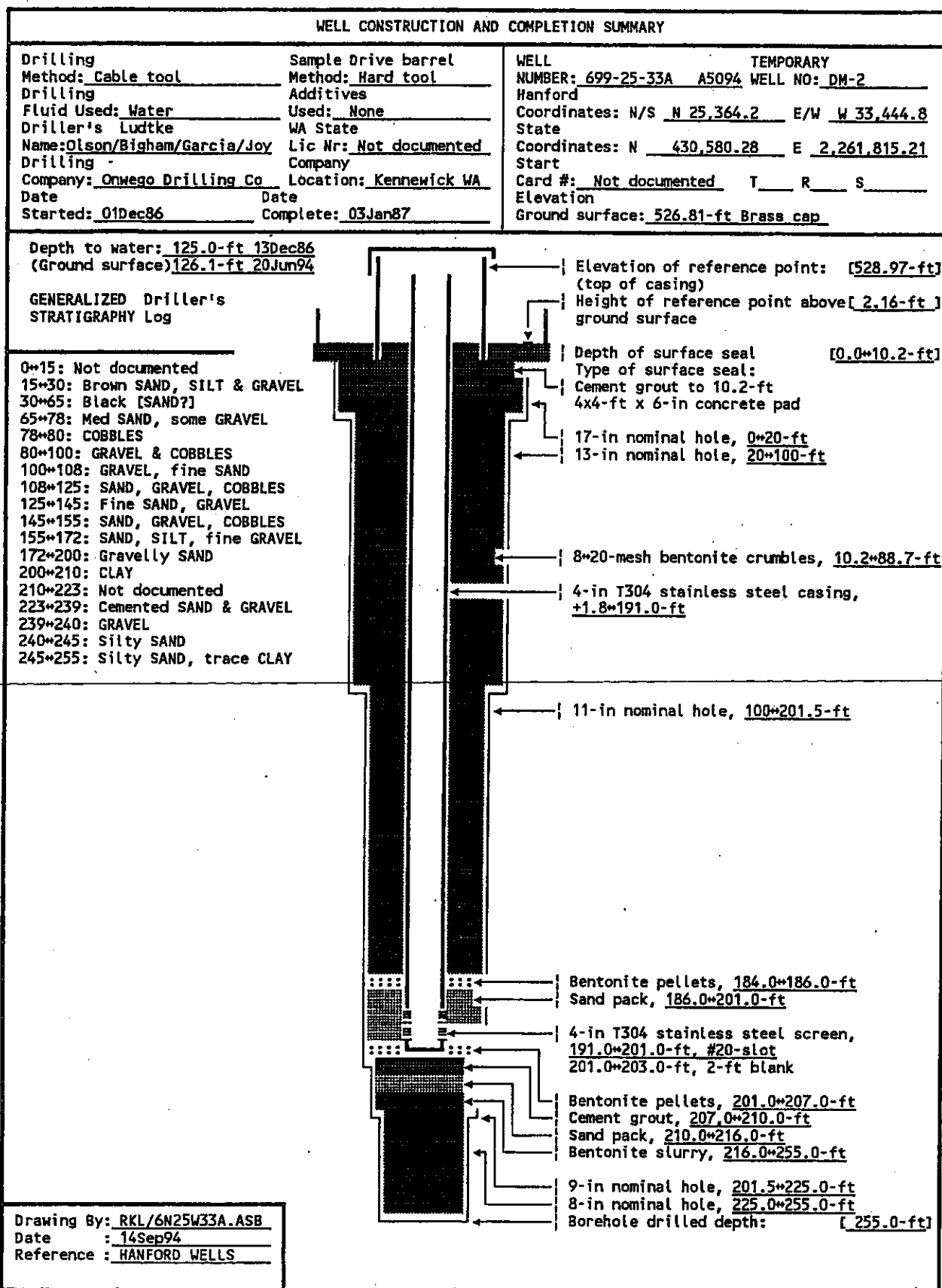
WAC 173-303, Washington Administrative Code. *Dangerous waste regulations*. Olympia, Washington.

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Appendix A

Construction Details of the Wells in the Monitoring Network



WELL CONSTRUCTION AND COMPLETION SUMMARY

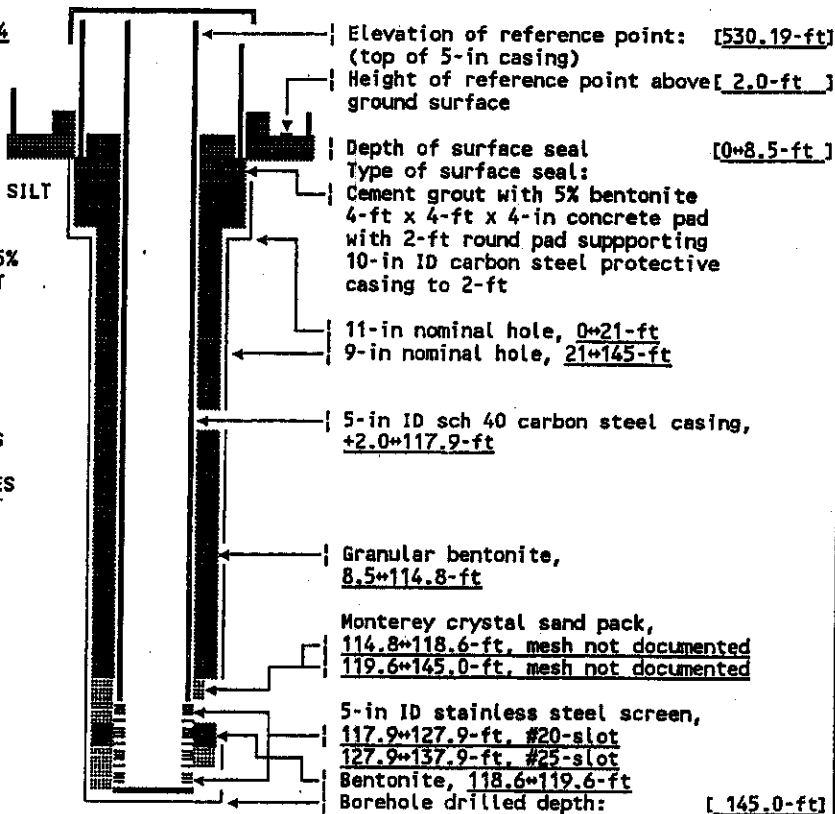
Drilling Method: Cable tool
 Drilling Fluid Used: Water
 Driller's Name: Evans/Myrick
 Company: Myrick's Well Drill
 Date Started: 20May86
 Sample Drive barrel Method: Hard tool
 Additives: None documented
 WA State Lic Nr: Not documented
 Location: Not documented
 Date Complete: 14Jul86

WELL NUMBER: 699-25-34A A5045 TEMPORARY WELL NO: SM-3
 Hanford
 Coordinates: N/S N 25,384.01 E/W W 33,500.86
 State
 Coordinates: N 430,599.95 E 2,261,759.09
 Start
 Card #: Not documented T R S
 Elevation
 Ground surface: 528.23-ft Brass cap

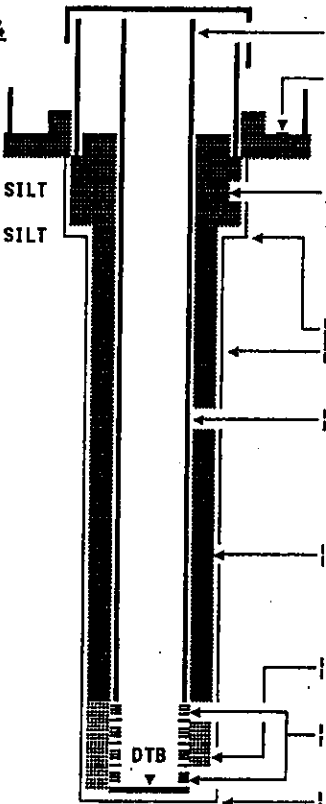
Depth to water: 125.3-ft Jun86
 (Ground surface) 127.2-ft 20Jun94

GENERALIZED Driller's STRATIGRAPHY Log

0*10: Not documented
 10*21: 90% coarse*fine SAND, 10% SILT
 21*30: Not documented
 30*70: Black SAND
 70*85: 40% coarse*medium SAND, 45% PEBBLES & COBBLES, 5% SILT
 85*95: 45% coarse*fine SAND, 45% COBBLES, 10% SILT
 95*100: Coarse*fine SAND, SILT, PEBBLES & COBBLES
 100*105: Coarse SAND, with large PEBBLES and COBBLES
 105*111: Coarse*fine SAND, PEBBLES COBBLES & SILT
 111*122: Coarse*fine SAND, PEBBLES & COBBLES
 122*140: BOULDERS & GRAVEL
 140*145: Black SAND, some GRAVEL



Drawing By: RKL/6N25434A.ASB
 Date: 14Sep94
 Reference: HANFORD WELLS

WELL CONSTRUCTION AND COMPLETION SUMMARY			
Drilling Method: <u>Cable tool</u> Drilling Fluid Used: <u>Water</u> Driller's Name: <u>Evans/Myrick</u> Drilling Company: <u>Myrick's Well Drill.</u> Location: <u>Not documented</u> Date Started: <u>30Jul86</u> Date Complete: <u>05Sep86</u>	Sample Drive barrel Method: <u>Hard tool</u> Additives Used: <u>None documented</u> WA State Lic Nr: <u>Not documented</u> Company	WELL NUMBER: <u>699-25-348</u> <u>A5096</u> TEMPORARY WELL NO: <u>SM-5</u> Hanford Coordinates: N/S <u>N 25,221.61</u> E/W <u>W 33,551.98</u> State Coordinates: N <u>430,437.42</u> E <u>2,261,708.39</u> Start Card #: <u>Not documented</u> T <u> </u> R <u> </u> S <u> </u> Elevation Ground surface: <u>526.92-ft Brass cap</u>	
Depth to water: <u>123.3-ft Aug86</u> (Ground surface) <u>126.4-ft 20Jun94</u>			
GENERALIZED Driller's STRATIGRAPHY Log			
0*35: Not documented 35*50: 90% fine*coarse SAND, 10% SILT 50*58: 60% SILT, 40% SAND 58*63: 90% coarse*fine SAND, 10% SILT 63*67.5: Very fine*coarse SAND 67.5*75: 90% fine*very fine SAND, 10% SILT 75*85: GRAVEL, PEBBLES, COBBLES SAND & SILT 85*139.3: Not documented	 <div style="position: absolute; top: 220px; right: 50px; width: 60%; font-size: 0.8em;"> Elevation of reference point: <u>[529.31-ft]</u> (top of 5-in casing) Height of reference point above <u>[2.4-ft]</u> ground surface Depth of surface seal <u>[0-20-ft]</u> Type of surface seal: Cement grout with 5% bentonite 4-ft x 4-ft x 4-in concrete pad with 2-ft round pad supporting 10-in ID carbon steel protective casing to 2-ft 11-in nominal hole, <u>0-20-ft</u> 9-in nominal hole, <u>20-139.3-ft</u> 5-in ID sch 40 carbon steel casing, <u>+2.4-118.2-ft</u> Granular bentonite, <u>20-120.5-ft</u> Monterey crystal sand pack, <u>120.5-139.3-ft, mesh not documented</u> 5-in stainless steel screen, <u>118.2-138.2-ft, #20-slot</u> <u>128.2-138.2-ft, #25-slot</u> Borehole drilled depth: <u>[139.3-ft]</u> DTB=Depth to bottom, <u>136.8-ft, 25Jun91</u> </div>		
Drawing By: <u>RKL/6N25W348.AS8</u> Date : <u>14Sep94</u> Reference : <u>HANFORD WELLS</u>			

WELL CONSTRUCTION AND COMPLETION SUMMARY

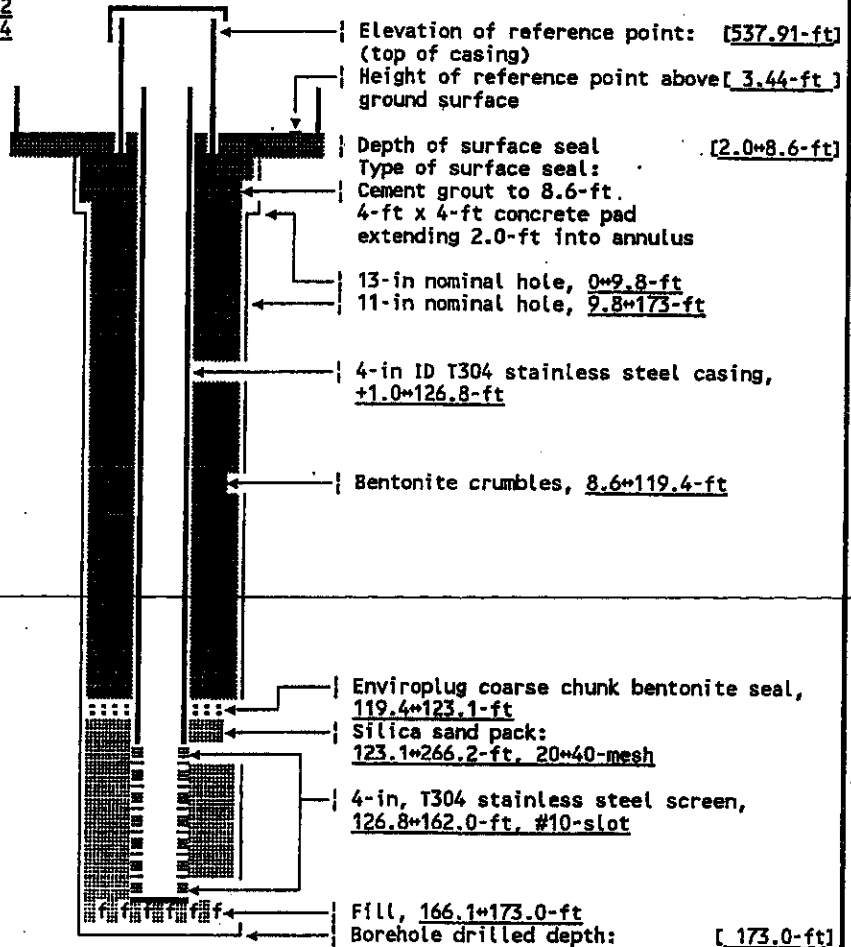
Drilling Backhoe 0-10-ft
 Method: ODEX air rotary
 Drilling Additives
 Fluid Used: None
 Driller's
 Name: S. McKinnon/D. Mingo
 Company: Jensen Drilling Co
 Date
 Started: 09Sep92
 Sample Method: Air returns
 Additives
 Used: None
 WA State
 Lic Nr: Not documented
 Company
 Location: Not documented
 Date
 Complete: 22Oct92

WELL TEMPORARY
 NUMBER: 699-25-34D A5419 WELL NO:
 Hanford
 Coordinates: N/S N 25,253.6 E/W W 33,847.9
 State NAD83 N 131,191.16m E 579,589.97m
 Coordinates: N 430,469 E 2,261,412
 Start
 Card #: Not documented T R S
 Elevation
 Ground surface: 534.47-ft (Brass cap)

Depth to water: 132.4-ft 22Oct92
 (Ground surface) 133.3-ft 20Jun94

GENERALIZED Geologist's
 STRATIGRAPHY Log
 Sl=slightly

0-84: SAND
 84-101: GRAVEL
 101-103: SAND
 103-126: GRAVEL
 126-126.5: Silty sandy GRAVEL
 126.5-161: Sandy GRAVEL
 161-173: Silty sandy GRAVEL



Drawing By: RKL/6N25W34D.ASB
 Date : 14Sep94
 Reference : WHC-SD-EN-DP-055

WELL CONSTRUCTION AND COMPLETION SUMMARY		
Drilling Method: <u>Cable tool</u> Fluid Used: <u>Water</u> Driller's Name: <u>Myrick</u> Company: <u>Myrick's Well Drill</u> Date Started: <u>01Aug86</u>	Sample Drive barrel Method: <u>Hard tool</u> Additives Used: <u>Bentonite</u> WA State Lic Nr: <u>Not documented</u> Company Location: <u>Not documented</u> Date Complete: <u>04Sep86</u>	WELL NUMBER: <u>699-26-33</u> TEMPORARY A5101 WELL NO: <u>SM-4</u> Hanford Coordinates: N/S <u>N 25,545.85</u> E/W <u>W 33,451.61</u> State Coordinates: N <u>430,761.92</u> E <u>2,261,807.92</u> Start Card #: <u>Not documented</u> T <u> </u> R <u> </u> S <u> </u> Elevation Ground surface: <u>533.75-ft Brass cap</u>
Depth to water: <u>128.5-ft Aug86</u> (Ground surface) <u>132.9-ft 20Jun94</u>		
GENERALIZED Driller's STRATIGRAPHY Log		
0*2: River run 2*20: Fine brown SAND, damp 20*25: Coarse black SAND, GRAVEL, PEBBLES, damp 25*28: Fine brown SAND, damp 28*30: SAND, dark, coarse 30*37.5: Coarse brown & black SAND, 1/4-in GRAVEL 37.5*45: BOULDER 45*45.5: Compact SAND 45.5*100: No record, drilled with drive barrel for moisture samples 100*108: Large GRAVEL 108*112: BOULDER 112*130: Large GRAVEL 130*135: SAND and GRAVEL 135*145: Very fine brown SAND 145*147: Fine SAND		Elevation of reference point: <u>[535.66-ft]</u> (top of 5-in casing) Height of reference point above <u>[1.9-ft]</u> ground surface Depth of surface seal <u>[0*20-ft]</u> Type of surface seal: Cement grout with 5% bentonite to 20-ft 4-ft x 4-ft x 4-in concrete pad with 2-ft round pad supporting 10-in ID carbon steel protective casing to 2-ft 11-in nominal hole, <u>0*20-ft</u> 9-in nominal hole, <u>20*147-ft</u> 5-in ID sch 40 carbon steel casing, <u>+1.9*123.5-ft</u> Granular bentonite, <u>20*120.5-ft</u> Monterey crystal sand pack, <u>120.5*147.0-ft</u> mesh not documented 5-in stainless steel screen, <u>123.5*143.5-ft, #25-slot</u> Borehole drilled depth: <u>[147.0-ft]</u> DTB=Depth to bottom, <u>143.7-ft, 25Jun91</u>
Drawing By: <u>RKL/6N26W33.ASB</u> Date: <u>14Sep94</u> Reference: <u>HANFORD WELLS</u>		

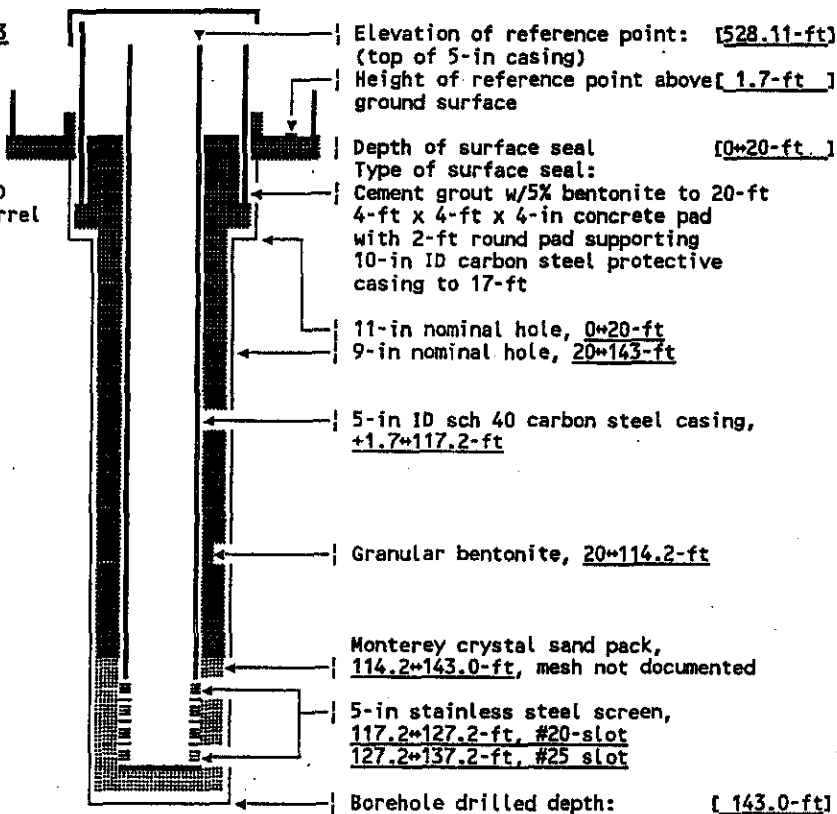
WELL CONSTRUCTION AND COMPLETION SUMMARY

Drilling Method: <u>Cable tool</u>	Sample Drive barrel Method: <u>Hard tool</u>	WELL NUMBER: <u>699-26-34A</u> A5102 WELL NO: <u>SM-2</u>
Drilling Fluid Used: <u>Water</u>	Additives Used: <u>None documented</u>	Hanford
Driller's Name: <u>Evans/Myrick(s)</u>	WA State Lic Nr: <u>Not documented</u>	Coordinates: N/S <u>N 26,162.95</u> E/W <u>W 34,483.26</u>
Drilling Company: <u>Myrick's Well Drill</u>	Location: <u>Not documented</u>	State
Date Started: <u>22May86</u>	Date Complete: <u>03Jul86</u>	Coordinates: N <u>431,376.37</u> E <u>2,260,774.69</u>
		Start
		Card #: <u>Not documented</u> T <u> </u> R <u> </u> S <u> </u>
		Elevation
		Ground surface: <u>526.43 Brass cap</u>

Depth to water: 121.5-ft Jun86
(Ground surface) 125.0-ft 27Dec93

GENERALIZED Driller's STRATIGRAPHY Log

0*75: Not documented
75*81: Large GRAVEL and fine SAND
81*120: Not documented, drive barrel
120*136: BOULDER or BOULDERS
136*143: SAND (heaving)



Drawing By: RKL/6N26W34A.ASB
Date: 14Sep94
Reference: HANFORD WELLS

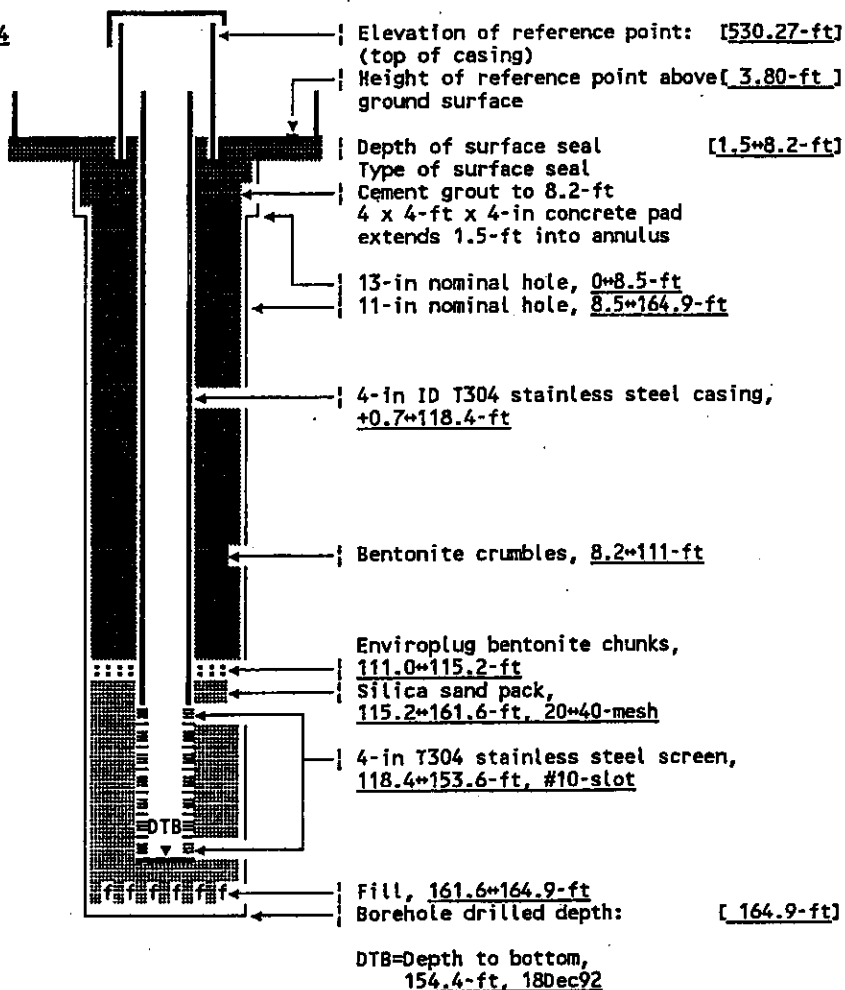
WELL CONSTRUCTION AND COMPLETION SUMMARY

Drilling Method: <u>Air rotary</u>	Sample Method: <u>Air returns</u>	WELL NUMBER: <u>699-26-34B</u>	TEMPORARY WELL NO: <u>A5420</u>
Drilling Fluid Used: <u>Not applicable</u>	Additives Used: <u>None</u>	Hanford	Coordinates: N/S <u>N 25,782.6</u> E/W <u>W 33,716.8</u>
Driller's Name: <u>S. McKennon/D. Mingo</u>	WA State Lic Nr: <u>Not documented</u>	State NAD83 N <u>131,352.50m</u> E <u>579,629.52m</u>	Coordinates: N <u>430,998</u> E <u>2,261,542</u>
Drilling Company: <u>Jensen</u>	Company Location: <u>Not documented</u>	Start Card #: <u>Not documented</u>	T <u> </u> R <u> </u> S <u> </u>
Date Started: <u>09Sep92</u>	Date Complete: <u>22Oct92</u>	Elevation Ground surface: <u>526.47</u>	Brass cap

Depth to water: 124.1-ft Oct92
(Ground surface) 125.3-ft 20Jun94

GENERALIZED Geologist's STRATIGRAPHY Log

0*74: SAND
74*82: GRAVEL
82*85: Sandy GRAVEL
85*87: GRAVEL
87*99: Gravelly SAND
99*102: GRAVEL
102*108: Sandy GRAVEL
108*119: SAND
119*122: Gravelly SAND
122*124: SAND
124*127: Sandy GRAVEL
127*139: Silty sandy GRAVEL
139*164.9: Pebble cobble GRAVEL



Drawing By: RKL/6N26W34B.ASB
Date : 14Sep94
Reference : WHC-SD-EN-DP-055

WELL CONSTRUCTION AND COMPLETION SUMMARY

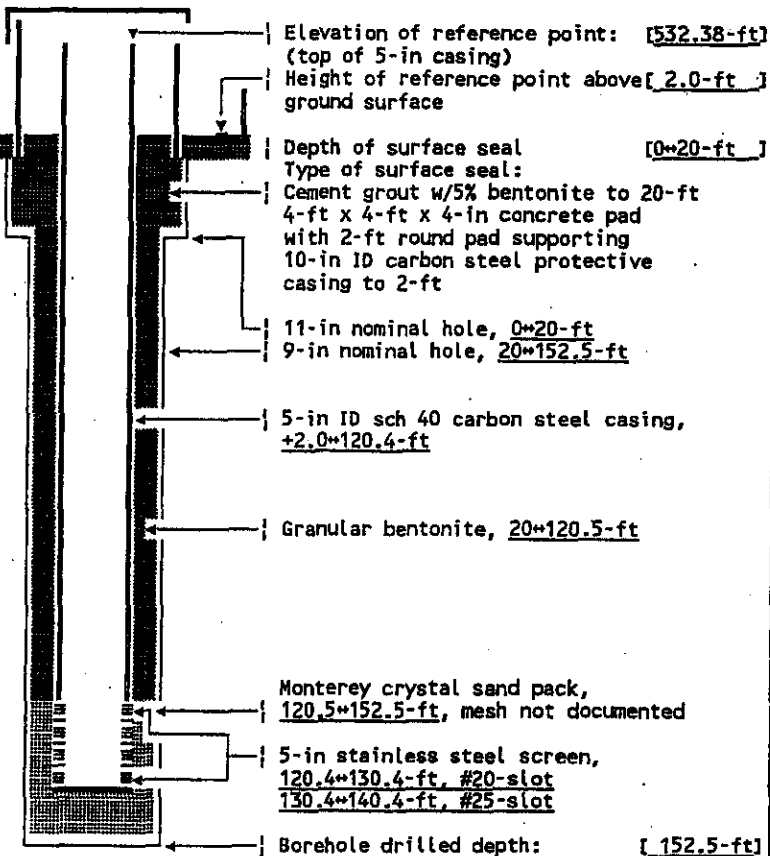
Drilling Method: Cable tool
 Drilling Fluid Used: Water
 Driller's Name: Myricks(s)
 Company: Myrick's Well Drill
 Date Started: 20May86
 Sample Drive barrel Method: Hard tool
 Additives Used: Bentonite
 WA State Lic Nr: Not documented
 Location: Not documented
 Date Complete: 14Jul86

WELL TEMPORARY
 NUMBER: 699-26-35A A5103 WELL NO: SM-1
 Hanford
 Coordinates: N/S N 25,768.80 E/W W 34,748.81
 State
 Coordinates: N 430,981.54 E 2,260,510.15
 Start
 Card #: Not documented T R S
 Elevation
 Ground surface: 530.38 Brass cap

Depth to water: 126.5-ft Jun86
 (Ground surface) 129.7-ft 20Jun94

GENERALIZED Driller's STRATIGRAPHY Log

0-38: Not documented
 38-69: 90% coarse SAND, 10% fine SAND
 69-70: 20% GRAVEL, 80% fine SAND
 70-95: Not documented
 95-114: BOULDERS
 114-120: GRAVEL
 120-130: GRAVEL & SAND
 130-152.5: SAND & GRAVEL



Drawing By: RKL/6N26W35A.ASB
 Date: 14Sep94
 Reference: HANFORD WELLS

WELL CONSTRUCTION AND COMPLETION SUMMARY		
Drilling Method: <u>Cable tool</u> Fluid Used: <u>Water</u> Driller <u>L Myrick/R Myrick</u> Name: <u>L Bultena</u> Drilling <u>Myrick Drilling Co</u> Company: <u>Orwego Drilling Co</u> Date Started: <u>23Jun86</u>	Sample Drive barrel Method: <u>Hard tool</u> Additives Used: <u>None</u> WA State Lic Nr: <u>Not documented</u> Company <u>Not documented</u> Location: <u>Kennewick WA</u> Date Complete: <u>05Jan87</u>	WELL NUMBER: <u>699-26-35C</u> TEMPORARY <u>A5104</u> WELL NO: <u>DM-1</u> Hanford Coordinates: N/S <u>N 25,751.10</u> E/W <u>W 34,688.94</u> State Coordinates: N <u>430,981.54</u> E <u>2,260,510.15</u> Start Card #: <u>Not documented</u> T <u> </u> R <u> </u> S <u> </u> Elevation Ground surface: <u>530.65-ft</u> Brass cap
Depth to water: <u>127.0-ft</u> 15Jul86 (Ground surface) <u>129.5-ft</u> 20Jun94		
GENERALIZED Driller's STRATIGRAPHY Log 0*4: Very cse b&w SAND 15*20: Fine brown SAND 20*30: Cse black SAND 30*40: Fine brown & black SAND 40*50: Black cse SAND 50*55: Fine grey SAND 55*65: Fine brown SAND 65*75: Black & brown SAND, w/ROCK 75*80: Brown SAND 80*85: Large ROCK, some SAND 85*90: Round river run ROCK 90*100: BOULDERS, GRAVEL, SAND 100*110: Round ROCK 110*115: ROCK, GRAVEL & SAND 115*125: BOULDERS 125*135: Not documented 135*140: Fine SAND 140*145: Fine compacted SAND 145*155: Fine SAND, heaving 155*170: Cse GRAVEL-SAND 170*172: Loose COBBLES 172*203: GRAVEL, SAND, some SILT 203*204: GRAVEL, SAND, some SILT, CLAY 204*211: Not documented (Attempted to sample w/drive sampler and split spoon)		
Drawing By: <u>RKL/6N26W35C.ASB</u> Date : <u>14Sep94</u> Reference : <u>HANFORD WELLS</u>		

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